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## An update of the assessment of the eastern stock of tarakihi for 2019

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A.D. Langley

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## EXECUTIVE SUMMARY

## Langley, A.D. (2019). An update of the assessment of the eastern stock of tarakihi for 2019.

New Zealand Fisheries Assessment Report 2019/41. 29 p.
Tarakihi (Nemadactylus macropterus) off the east coast of the North and South Island are considered to represent a single biological stock. The domain of the stock encompasses the TAR 3 and TAR 2 Fishstocks, the eastern portion of the TAR 1 Fishstock and a small area of the TAR 7 Fishstock in eastern Cook Strait. The first assessment of the eastern tarakihi stock was completed in November 2017. The stock assessment was conducted using a statistical, age-structured population model implemented in Stock Synthesis.
The assessment model is structured with a single region and four fisheries and commences in 1975 under exploited, equilibrium conditions. The 2017 assessment model was updated with the additional data available from the 2016/17 and 2017/18 fishing years, specifically two years of annual catch, two additional CPUE indices and the biomass estimate and length composition from the 2017 ECSI Kaharoa trawl survey. Annual catches from the most recent two years (2017 and 2018) were generally comparable to the three previous years (2014-2016), with the exception of the TAR 3 trawl fishery. Catches from that fishery increased by $45 \%$ from 2014 to 2017 and then dropped by $33 \%$ in 2018.

Five sets of CPUE indices were updated based on the previous CPUE analyses. For the TAR 3 trawl fishery, the CPUE indices increased during 2013/14-2016/17 and then dropped in 2017/18, following the trend in catch from the fishery. The CPUE indices from the TAR 3 set net fishery increased during 2014/15-2017/18. Annual combined CPUE indices from the TAR 2 trawl fishery increased by $51 \%$ from $2013 / 14$ to $2016 / 17$ and then dropped by $14 \%$ in $2017 / 18$. The CPUE indices from the Bay of Plenty trawl fishery were at the lowest level of the time series during 2013/14-2015/16, increased by $37 \%$ in 2016/17 and remained at a similar level in 2017/18. The CPUE indices from the east Northland trawl fishery declined by $60 \%$ during 2009/10-2017/18 with a $33 \%$ drop in 2017/18.

The updated model provided a very good fit to the recent CPUE indices from the TAR2/Bay of Plenty trawl fishery and the Kaikoura set net fishery. However, there were a number of discrepancies in the fits to the other recent observations, especially the CPUE indices from the East Northland and TAR 3 trawl fisheries and the recent ECSI trawl survey biomass estimates.

The updated assessment model indicates that there was no appreciable change in stock status from 2016 $(2015 / 16)$ to $2018(2017 / 18)$. Current (2018) stock status was estimated to be at 0.159 SB $_{0}$ (C.I. 0.113$0.205)$. There is a very high probability ( $96 \%$ ) that the stock was below the soft limit and a negligible probability ( $<1 \%$ ) that the stock was below the $10 \% S B_{0}$ hard limit.

Forward projections of the updated assessment model were conducted at fractions ( $50 \%, 60 \%, 70 \%$, $80 \%, 90 \%$ and $100 \%$ ) of the base level of the model fishery catches (total $3560 t$ ) for the years 2020 to 2048. For each catch scenario, the median spawning biomass increased from the current (2018) level, although the rate of increase varied depending on the magnitude of the projected catch. There is a corresponding increase in the probability of the stock increasing above the $20 \% S B_{0}$ soft limit with the different levels of catch. The projections were also evaluated relative to two potential levels of target biomass for the stock ( $35 \%$ and $40 \% S B_{0}$ ).

A simulation analysis was conducted to investigate the performance of alternative target biomass levels for the stock. Target biomass levels were evaluated within the framework of a set of Harvest Control Rules (HCR) and the associated Management Procedures. The results are intended to assist managers in the formulation of an appropriate level of target biomass for the eastern tarakihi stock. The current study indicates that a target biomass level of about $35 \% S B_{0}$ is sufficiently high to minimise the risk of breaching the soft and hard limits, while maintaining catches at a relatively high level. This is somewhat lower than the current default target biomass level of $40 \% S B_{0}$. These results are predicated on the presumptions related to the management procedures and HCRs adopted for the simulations.

## 1. INTRODUCTION

Tarakihi (Nemadactylus macropterus) off the east coast of the North and South Island are considered to represent a single biological stock (Langley 2018). The domain of the stock encompasses the TAR 3 and TAR 2 Fishstocks, the eastern area of the TAR 1 Fishstock and a small area of the TAR 7 Fishstock in eastern Cook Strait.

The first assessment of the eastern tarakihi stock was completed in November 2017. The stock assessment was conducted using a statistical, age-structured population model implemented in Stock Synthesis (Methot \& Wetzell 2013). The assessment incorporated the available catch, CPUE indices, trawl survey biomass estimates and age/length frequency distributions, and recent commercial age compositions. The model data sets were structured into three areas: east coast South Island (including eastern Cook Strait), central east coast North Island and the Bay of Plenty combined (BPLE-TAR2), and East Northland. The east coast South Island area included three commercial fisheries: the Canterbury Bight/Pegasus Bay trawl fishery, Kaikoura set net fishery and the eastern Cook Strait trawl fishery. The other two areas included a commercial trawl fishery and a relatively small non-commercial fishery. For each area, a corresponding time-series (or multiple series) of CPUE indices was available (Langley 2017).
Spawning biomass was estimated to have been reduced to $22 \% S B_{0}$ by the mid 1970 s, following a period of relatively high catches (5000-7000 t) during the 1950s and early 1960s. For the base case assessment model, current (2015/16) spawning biomass was estimated to be at $17 \%$ of the unfished, equilibrium biomass level $\left(S B_{2016} / S B_{0}=0.170\right)$, which is below the soft limit of $20 \% S B_{0}$ (Langley 2018).

The stock assessment model was updated in early 2018 with the inclusion of catches and CPUE indices from the 2016/17 fishing year. The assessment update was funded by the commercial stakeholder groups Fisheries Inshore New Zealand (FINZ) and Southern Inshore Fisheries Management Company (SIFMC) and reported to the Southern Inshore Stock Assessment Working Group on 4 April 2018. The updated base case model yielded results that were consistent with the previous assessment and estimated the 2016/17 spawning biomass to be at $17 \%$ of the unfished, equilibrium biomass level $\left(S B_{2017} / S B 0=\right.$ 0.173 ). There was a high probability ( $87 \%$ ) that the spawning biomass was below the soft limit ( $20 \%$ $S B_{0}$ ) but a very low probability (less than $1 \%$ ) of being below the hard limit of $10 \% S B_{0}$.
The updated stock assessment model was applied to conduct a range of stock projections with different multiples of fishery catch. These projections were utilised in the formulation of management advice to implement a rebuild strategy for the stock. The management advice lead to reductions in the TACCs for TAR 1 (a $24 \%$ reduction), TAR 2 ( $16 \%$ ), TAR 3 ( $26 \%$ ) and TAR 7 ( $4 \%$ ) for the 2018/19 fishing year. A range of voluntary measures were also introduced in 2018/19, including measures to ensure that the reductions in the TAR 1 and TAR 7 TACCs were applied to the east coast portions of those Fishstocks.

The commercial stakeholder groups FINZ and SIFMC funded a further update of the stock assessment model in 2019, incorporating catch, CPUE and trawl survey data from the 2017/18 fishing year. The updated assessment model was applied to conduct an additional set of stock projections for consideration in the development of management advice for the 2019/20 fishing year. In addition, the stock assessment model was used in a preliminary evaluation of potential target biomass reference points for the eastern tarakihi stock.

## 2. STOCK ASSESSMENT INPUTS

The recent updates of the 2017 stock assessment have retained the equivalent model structure and data configuration. The most recent (2019) update of the assessment model incorporated two additional years of fishery catch data and CPUE indices and an additional set of data from the time series of winter east coast South Island (ECSI) Kaharoa trawl surveys (April-June 2018).

### 2.1 Fishery catches

The stock assessment model incorporates six commercial fisheries: a set net fishery off Kaikoura (TAR3-SN) and trawl fisheries in Canterbury Bight/Pegasus Bay (TAR3-BT), eastern Cook Strait
(Cook-BT), east coast North Island (TAR2-BT), Bay of Plenty (BPLE-BT) and east Northland (ENLDBT). The configuration of these fisheries is detailed in Langley (2018). Annual catches were compiled by fishing year; years in the assessment model are denoted by the calendar year at 1 January (e.g. the 2018 model year represents the 2017/18 fishing year).

Annual catches from the commercial fisheries for 2016/17 and 2017/18 fishing years were compiled from an extract of recent catch and effort data provided by Fisheries New Zealand (Data Extract 12270). These data were processed following the methodology described in Langley (2017). The fishery catches also included an additional allowance for unreported catches, assumed to represent $10 \%$ of the reported landings.

Annual catches from the most recent two years (2017 and 2018) were generally comparable to the three previous years (2014-2016), with the exception of the TAR3-BT fishery (Figure 1) which increased by $45 \%$ from 2014 to 2017 and then dropped by $33 \%$ in 2018.


Figure 1: Recent annual catches of tarakihi by commercial fishery and total (including recreational catch). Annual catches include allowances for unreported catch. Model years are configured by fishing year (denoted by the calendar year at 1 January).

The time series of annual fishery catches included in the assessment model is tabulated in Appendix 1. Catches from the recreational fisheries in 2017 and 2018 were assumed to be equivalent to those in the preceding years.

### 2.2 CPUE indices

For the 2017 stock assessment, standardised CPUE indices were derived for five tarakihi fisheries: TAR 3 trawl, TAR 3 set net, TAR 2 trawl, the Bay of Plenty trawl and east Northland trawl (Langley 2017). These sets of CPUE indices were updated in 2019 with the addition of two years of catch and effort data (2016/17 and 2017/18). The configuration of the individual CPUE data sets is described in Langley (2017).

The individual CPUE models were simply refitted with the equivalent set of explanatory variables included in the original analyses (Langley 2017). The updated CPUE models all yielded indices that were virtually identical to the corresponding annual indices from the original models.

For the TAR 3 trawl fishery, the combined (delta-lognormal) CPUE indices generally increased during 2013/14-2016/17 and then dropped in 2017/18, following the trend in catch from the fishery (Figure 2). The lognormal CPUE indices from the TAR 3 set net fishery increased during 2014/15-2017/18 (Figure 3). Annual combined CPUE indices from the TAR 2 trawl fishery increased by $51 \%$ from 2013/14 to 2016/17 and then dropped by $14 \%$ in 2017/18 (Figure 4).
The combined (delta-Weibull) CPUE indices from the Bay of Plenty trawl fishery were at the lowest level of the time series during 2013/14-2015/16 (Figure 5). The indices increased by $37 \%$ in 2016/17 and remained at a similar level in 2017/18. The combined (delta-Weibull) CPUE indices from the east Northland trawl fishery declined by $60 \%$ during 2009/10-2017/18 with a $33 \%$ drop in the CPUE index in 2017/18 (Figure 6).


Figure 2: Standardised CPUE indices from the lognormal, binomial and combined (delta-lognormal) CPUE models for the TAR 3 trawl fishery (the vertical lines represent the $\mathbf{9 5 \%}$ confidence intervals).


Figure 3: Standardised CPUE indices from the lognormal CPUE model for the TAR 3 set net fishery (the vertical lines represent the $\mathbf{9 5 \%}$ confidence intervals).


Figure 4: Standardised CPUE indices from the lognormal, binomial and combined (delta-lognormal) CPUE models for the TAR 2 trawl fishery (the vertical lines represent the $\mathbf{9 5 \%}$ confidence intervals).


Figure 5: Standardised CPUE indices from the Weibull, binomial and combined (delta-Weibull) CPUE models for the Bay of Plenty (TAR 1) trawl fishery (the vertical lines represent the $\mathbf{9 5 \%}$ confidence intervals).


Figure 6: Standardised CPUE indices from the Weibull, binomial and combined (delta-Weibull) CPUE models for the East Northland (TAR 1) trawl fishery (the vertical lines represent the $\mathbf{9 5 \%}$ confidence intervals).

For inclusion in the stock assessment model, the separate sets of CPUE indices from the TAR 2 trawl and Bay of Plenty trawl fisheries were combined (TAR2BPLE-BT), weighted by the annual catches from each fishery. The final sets of CPUE indices included in the updated stock assessment model are presented in Appendix 2. The four sets of CPUE indices were each assigned a coefficient of variation (CV) of $20 \%$ in the assessment model (Langley 2018).

The 2017 stock assessment model did not include standardised CPUE indices from the Cook Strait trawl fishery, primarily due to the relatively limited amount of catch and effort data available from the fishery and limited age composition data available to reliably estimate the selectivity of the fishery. However, during the update of the CPUE indices in 2018 the catch and effort data from the fishery (to 2016/17) were reanalysed and the resulting time-series of CPUE indices was accepted by the SINS WG.
The trends in the CPUE indices from the Cook Strait fishery (Figure 7) are generally consistent with the trends in the CPUE indices from the TAR 3 trawl and set net fisheries (Figure 2 and Figure 3) once the differences in the age composition of the catches between the three fisheries are taken into account (Langley 2018). These additional CPUE indices have not been incorporated in the subsequent updates of the stock assessment model.


Figure 7: Standardised CPUE indices from the lognormal, binomial and combined (delta-lognormal) CPUE models for the Cook Strait trawl fishery (the vertical lines represent the $\mathbf{9 5 \%}$ confidence intervals).

### 2.3 ECSI Kaharoa trawl survey

The stock assessment model includes the time-series of biomass estimates and length or age compositions from the ECSI Kaharoa trawl surveys. Since the completion of the stock assessment in 2017, there was an additional trawl survey conducted in April-June 2018 (MacGibbon et al. 2019). The tarakihi biomass estimate from the core area of the trawl survey was 1407 t (CV 0.25) which is lower than the geometric mean of the biomass estimates from the series of winter trawl surveys (1730 t). The trawl survey biomass is primarily composed of fish in the $1-5$ year age classes (Langley 2018).
Since 2007, age composition data have been available from each of the ECSI Kaharoa trawl surveys. However, the tarakihi otoliths that were collected during the 2018 trawl survey (MacGibbon et al. 2019) are yet to be aged and, in the interim, the 2018 survey length composition was included in the assessment model. The next iteration of the assessment model will instead incorporate the age composition from the latest survey.

## 3. STOCK ASSESSMENT MODEL

### 3.1 Model structure

The structure of the updated assessment model was essentially equivalent to the base-case stock assessment model completed in 2017 (Langley 2018). The model is structured with a single region and four fisheries with the age structure initialised in 1975 assuming exploited, equilibrium conditions. The updated model extended the model period from 2016 to 2018 and estimates current stock status in 2018 (2017/18 fishing year). The additional trawl survey observations (biomass and length composition) enable the time-series of recruitment deviates to be extended from 2015 to 2017.
The model was implemented in Stock Synthesis version 3.24 Z which was the same version of the software used in the original assessment. The model objective function included contributions from the fishery catches, initial equilibrium catches, indices of abundance (CPUE and survey), age-compositions (commercial and survey), length-compositions, (survey) recruitment, and priors and penalties (see Methot \& Wetzell 2013). The weighting applied to the individual data observations was equivalent to the original assessment model.

The estimation procedure minimises the negative log-likelihood of the objective function to determine the mode of the joint posterior distribution (MPD). Model uncertainty was determined using Markov chain Monte Carlo (MCMC) implemented using the Metropolis-Hastings algorithm; 1000 MCMC samples were drawn at 1000 intervals from a chain of 1.1 million following an initial burn-in of 100000 .

### 3.2 Model diagnostics

The model fit to the individual data sets was very similar to the fit to the comparative data sets in the 2017 stock assessment model. For the MPDs, the combined age composition likelihoods were very similar for the two models ( 33.84 compared to 33.23 from the 2017 assessment) and the models provide a good fit to the individual age compositions from the trawl surveys and fisheries (Langley 2018). The key parameter estimates were also very similar.

As with the original model, the updated assessment provides a good fit to the general trend in four sets of CPUE indices, although there is some divergence from each of the sets of CPUE indices over the time-series (Figure 8). For the most recent two years, the updated model provides a good fit to the TAR2BPLE-BT and TAR3-SN CPUE indices. However, the model does not fit the large drop in the ENLD-BT CPUE index in 2018 and underestimates the TAR3-BT CPUE indices in 2017 and 2018 (Figure 8).
The recent TAR3-BT CPUE indices appear to conflict with the biomass estimates from the two most recent ECSI trawl surveys; the model underestimates the CPUE indices but overestimates the trawl survey vulnerable biomass in both 2016 and 2018 (Figure 9). As in the original assessment, the model fits to the length compositions from the ECSI trawl survey are generally quite poor and do not adequately fit the modal structure of the individual cohorts (Figure 10). This observation also pertains to the length composition from the most recent (2018) trawl survey.


Figure 8: Fits to the four sets of CPUE indices included in the updated assessment model.


Figure 9: Observed (grey points) and predicted (blue triangles) winter ECSI trawl survey biomass from the updated assessment model.


Figure 10: Comparison of observed (points) and predicted (lines) individual length compositions (both sexes combined) from the winter ECSI Kaharoa trawl surveys (including the most recent survey) from the updated assessment model.

The updated model estimated annual recruitments that were very similar in magnitude to the original assessment (Figure 11). There was a small reduction in the level of recruitment estimated during the early 1980s and recruitment estimates from the terminal years of the original model (2013-2015) were less variable. Recent recruitments are characterised by relatively strong year classes in 2011 and 2012 and lower recruitment in 2013 and 2014. Recruitments in 2015-2017 were poorly estimated by the updated assessment model (Figure 11), reflecting the limited information from the recent ECSI trawl survey data.


Figure 11: A comparison of the estimates of annual recruitment from the 2017 assessment (red dashed line) and the updated assessment (black line and associated $95 \%$ confidence intervals). The lines represent the median values of the MCMC distributions.

### 3.3 Stock status

The updated assessment model yielded stock trajectories that were very similar to the original assessment model, in terms of absolute spawning biomass (Figure 12) and spawning biomass relative to equilibrium unexploited biomass (Figure 13). There was some divergence in the stock trajectories in the 1980s and early 1990s corresponding to the lower level of recruitment estimated by the updated assessment model (Figure 11). The estimate of stock status in 2016, the terminal year of the original assessment, was also slightly lower from the updated model $\left(S B_{2016} / S B_{0} 0.161\right.$ compared to 0.170 from the previous assessment) with a corresponding lower probability of being above the soft limit of $20 \%$ $S B_{0}$ (Table 1).

The updated assessment model indicates that there was no appreciable change in stock status from 2016 (2015/16) to 2018 (2017/18) (Table 1 and Table 2). Current (2018) stock status was estimated to be at $0.159 S B_{0}$ (C.I. $0.113-0.205$ ). There was a very high probability ( $96 \%$ ) that the stock was below the soft limit and a negligible probability ( $<1 \%$ ) that the stock was below the $10 \% S B_{0}$ hard limit (Table 2). The estimates of current stock status were consistent with the results of stock projections conducted with the original assessment model (based on an assumption of constant catches equivalent to the 2016 catch levels) (Table 2).


Figure 12: A comparison of spawning biomass from the 2017 assessment (red dashed line) and the updated assessment (black line and associated $\mathbf{9 5 \%}$ confidence intervals). The lines represent the median values of the MCMC distributions.


Figure 13: A comparison of spawning biomass relative to virgin biomass ( $S B_{0}$ ) from the 2017 assessment (red dashed line) and the updated assessment (black line and associated $95 \%$ confidence intervals). The lines represent the median values of the MCMC distributions.

Table 1: Stock status in 2016 (2015/16 fishing year) from the original assessment (Base 2017) and the updated assessment model.

| Model option | $S B_{0}$ | SB 2018 | $S^{\text {B }}$ 2018/ $/$ SB $_{0}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 40\% | 20\% | 10\% |
| Base 2017 | $\begin{gathered} 86321 \\ (81977-91907) \end{gathered}$ | $\begin{gathered} 14620 \\ (10685-19413) \end{gathered}$ | $\begin{gathered} 0.170 \\ (0.126-0.219) \end{gathered}$ | 0.00 | 0.112 | 0.997 |
| Update 2019 | $\begin{gathered} 86972 \\ (82432-92164) \end{gathered}$ | $\begin{gathered} 13955 \\ (10330-17985) \end{gathered}$ | $\begin{gathered} 0.161 \\ (0.120-0.204) \end{gathered}$ | 0.00 | 0.036 | 0.998 |

Table 2: Stock status in 2018 (2017/18 fishing year) from the original assessment (Base 2017) and the updated assessment model. The stock status from the original assessment is based on a constant catch projection, assuming projected catches were equivalent to 2016 catches. Projection results are presented in grey italics.

| Model option | $\boldsymbol{S B} \mathbf{O}_{0}$ | SB2018 | $\mathrm{SB}_{2018} / \mathrm{SB}_{0}$ | $\operatorname{Pr}\left(\mathrm{SB}_{2018}>\boldsymbol{X} \% \mathrm{SB}_{0}\right)$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 40\% | 20\% | 10\% |
| Base 2017 | $\begin{gathered} 86321 \\ (81 \text { 977-91 907) } \end{gathered}$ | $\begin{gathered} 13671 \\ (8002-19900) \end{gathered}$ | $\begin{gathered} 0.158 \\ (0.098-0.222) \end{gathered}$ | 0.00 | 0.148 | 0.998 |
| Update 2019 | $\begin{gathered} 86972 \\ (82432-92164) \end{gathered}$ | $\begin{gathered} 13844 \\ (9762-18220) \end{gathered}$ | $\begin{gathered} 0.159 \\ (0.113-0.205) \end{gathered}$ | 0.00 | 0.042 | 0.996 |

### 3.4 Forward projections

The updated assessment model was applied to conduct forward projections to evaluate a range of different catch scenarios, specifically related to the rebuilding of the stock. Forward projections were conducted for a 30 year period from 2019 to 2048 . For all scenarios, annual catches in the first year (2019 = 2018/19) were equivalent at the base level of catches corresponding to the new TACCs introduced in 2018/19 and the associated proportion allocated to the eastern stock (for TAR 1 and TAR 7) (Table 3).

The 2018/19 catches for each Fishstock/QMA were then apportioned between model fisheries based on the catch proportions from 2017/18 (Table 3). The Cook Strait trawl fishery operates within TAR 7 and the northern area of TAR 3 (Statistical Area 018), and hence the Cook-BT fishery catch is composed of a proportion of the TACC of both Fishstocks. The projected commercial catches also included an additional $10 \%$ unreported catch (Table 3). Recreational catches were held constant throughout the projection period at the 2016 level ( 71 t and 97 t from Bay of Plenty and QMA 2, respectively).

Forward projections were conducted at fractions of the base level of the model fishery catches (total 3560 t) for the years 2020 to 2048. The projections were conducted for $50 \%, 60 \%, 70 \%, 80 \%, 90 \%$ and $100 \%$ of the base levels of fishery catch.

Table 3: Derivation of the base level of fishery catches included in the forward projections.

| Fishstock | $\begin{array}{r} \text { 2018/19 } \\ \text { TACC } \end{array}$ | Proportion East | $\begin{array}{r} 2018 / 19 \\ \text { Catch } \end{array}$ | Model fishery (\% 2018/19 catch) | Catch (t) | Including 10\% unreporting |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TAR 1 | 1097 | 0.472 | 518 | ENLD-BT (45\%) | 233 | 256 |
|  |  |  |  | BPLE-BT (55\%) | 285 | 313 |
| TAR 2 | 1500 | 1.0 | 1500 | TAR2-BT (100\%) | 1500 | 1650 |
| TAR 3 | 1040 | 1.0 | 1040 | TAR3-BT (66\%) | 688 | 757 |
|  |  |  |  | TAR3-SN (10\%) | 106 | 116 |
|  |  |  |  | Cook-BT (24\%) | 246 | 67 |
| TAR 7 | 1042 | 0.17 | 179 | Cook-BT (100\%) | 179 | \% |
| Total |  |  |  |  | 3237 | 3560 |

Forward projections were conducted using MCMC with 1000 samples drawn at 1000 intervals from a chain of 1.1 million draws (with an initial burn-in of 100000 ). Annual recruitments were derived from the Beverton-Holt spawner-recruit relationship with recruitment deviates resampled from the assumed distribution (sigmaR 0.6).

The results of the individual projections were collated to determine the median biomass level $\left(S B_{\text {year }} / S B_{0}\right)$ and the probability of the biomass being above the hard limit ( $10 \% S B_{0}$ ), the soft limit ( $20 \%$ $S B_{0}$ ) and two potential levels of target biomass ( $35 \%$ and $40 \% S B_{0}$ ) for each year of the projection. The full range of results from the projections is presented in Appendix 3.

For each catch scenario, median spawning biomass increases from the current (2018) level, although the rate of increase varies depending on the magnitude of the projected catch (Figure 14). For example, under the current $(2018 / 19)$ level of commercial catch, the median biomass increases from 0.158 to 0.23 $S B / S B_{0}$ in 10 years, whereas at $50 \%$ of the current catch level the stock reaches $0.40 S B / S B_{0}$ in the same period.


Figure 14: Median annual spawning biomass relative to virgin spawning biomass ( $\mathrm{SB}_{0}$ ) for the projection period for each of the catch scenarios (percentages of the base level of catch).

There is a corresponding increase in the probability of the stock increasing above the $20 \% S B_{0}$ soft limit with the different levels of catch. The current ( $100 \%$ ) catch projections indicate that there is a $62 \%$ probability that the stock will be above the soft limit in 10 years, whereas there is a $99 \%$ probability of being above the soft limit in 10 years with catches at $50 \%$ of the base level (Figure 15).

The probability of reaching $35 \%$ and $40 \% S B_{0}$ increases over time for all catch scenarios, although the probability of reaching either threshold during the projection period is higher for the lower catch scenarios (Figure 15).
For all catch scenarios, the projections indicate that the risk of the stock declining below the $10 \%$ hard limit is negligible over the next 10 years. The probability remains very low for all catch scenarios throughout the projection period, with the exception of the base catch scenario $(100 \%)$ which has a 5$10 \%$ probability of decreasing below the hard limit throughout the last 20 years of the projection period (Appendix 3).




$$
\begin{array}{|ll|}
- & \text { Catch } 50 \% \\
- & \text { Catch } 60 \% \\
- & \text { Catch } 70 \% \\
- & \text { Catch } 80 \% \\
\text { Catch } 90 \% \\
& \text { Catch } 100 \% \\
\hline
\end{array}
$$

Figure 15: Annual probability of spawning biomass being greater than $20 \%$ (bottom left panel), $\mathbf{4 0 \%}$ (top left) and $35 \%$ (top right) $S B_{0}$ for the projection period for each of the catch scenarios (percentages of the base level of catch).

## 4. DISCUSSION

The 2017 assessment model was updated with the additional data available from 2016/17 and 2017/18 fishing years. The updated model provided a very good fit to the recent CPUE indices from the TAR2/Bay of Plenty trawl fishery and Kaikoura set net fishery. However, there are a number of discrepancies noted in the fits to the other recent observations, especially the CPUE indices from the East Northland and Canterbury Bight/Pegasus Bay trawl fisheries and the recent ECSI trawl survey biomass estimates. The deterioration in the fit to these data may indicate regional differences in recent trends in stock abundance due to differential levels of exploitation and/or differences in recent patterns in recruitment. A national catch sampling programme has been implemented to collect age composition data from the main tarakihi fisheries during the 2018/19 and 2019/20 fishing years, and these data will be available for incorporation into the next full stock assessment of eastern tarakihi, planned for 2021. Age composition data will also be available from the 2018 and 2020 ECSI trawl survey. The age composition data are likely to increase our understanding of the regional scale stock dynamics and assist in elucidating recent trends in stock biomass.

The relatively poor fit to the most recent abundance indices indicates that further evaluation of the reliability of the CPUE indices from the main fisheries is required. The recent changes in the management of the tarakihi fishery (i.e. the reductions in TACCs and the partitioning of catches within the TAR 1 Fishstock) are likely to influence the operation of the trawl fisheries which may influence the reliability of the CPUE indices, especially in the coming years.

The relatively poor fit to some of the most recent abundance indices introduces additional uncertainty into the estimates of current stock status, although the overall assessment model was relatively insensitive to the inclusion of these data. Therefore, it is considered that the updated assessment model is sufficiently robust for the purposes of conducting forward projections. The current set of forward projections yielded results that are entirely consistent with projections conducted with an earlier (2018) iteration of the model.

## 5. MANAGEMENT IMPLICATIONS

The results of the forward projections have been provided to Fisheries New Zealand and commercial stakeholder groups (FINZ and SIFMC) for the formulation of management proposals for consideration in the 2019 Sustainability Round.

The next full stock assessment for eastern tarakihi is scheduled for 2021 and will incorporate the age compositions of the catch derived from sampling during 2018/19 and 2019/20 and from the 2018 and 2020 ECSI trawl surveys. It is anticipated that the updated stock assessment will improve estimates of current stock status and enable the initial period of the rebuilding of the stock to be evaluated (to 2020/21).

## 6. ACKNOWLEDGMENTS

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## APPENDIX 1 MODEL CATCH HISTORY

Table A1: Annual tarakihi catch (t) by fishery and total included in the stock assessment model.

| Year |  |  |  |  |  |  | Fishery |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{array}{r} \text { TAR3- } \\ \text { BT } \end{array}$ | $\begin{array}{r} \text { TAR3- } \\ \text { SN } \end{array}$ | Cook-BT | TAR2- BT | TAR2- <br> Rec | $\begin{array}{r} \text { BPLE- } \\ \text { BT } \end{array}$ | BPLE- <br> Rec | $\begin{array}{r} \hline \text { ENLD- } \\ \text { BT } \end{array}$ |  |
| 1975 | 2296 | 0 | 498 | 1769 | 0 | 426 | 0 | 231 | 5220 |
| 1976 | 1327 | 0 | 449 | 1643 | 0 | 260 | 0 | 327 | 4006 |
| 1977 | 2312 | 0 | 590 | 1384 | 0 | 385 | 0 | 231 | 4902 |
| 1978 | 1978 | 0 | 744 | 1649 | 0 | 600 | 0 | 205 | 5176 |
| 1979 | 448 | 0 | 559 | 1502 | 0 | 797 | 0 | 347 | 3653 |
| 1980 | 816 | 43 | 500 | 1150 | 0 | 735 | 0 | 503 | 3746 |
| 1981 | 1186 | 132 | 504 | 1164 | 71 | 805 | 97 | 603 | 4561 |
| 1982 | 1266 | 223 | 394 | 1213 | 71 | 594 | 97 | 553 | 4411 |
| 1983 | 915 | 229 | 216 | 1316 | 71 | 743 | 97 | 433 | 4020 |
| 1984 | 866 | 216 | 362 | 979 | 71 | 730 | 97 | 554 | 3877 |
| 1985 | 1232 | 308 | 384 | 971 | 71 | 615 | 97 | 448 | 4125 |
| 1986 | 1101 | 275 | 486 | 1096 | 71 | 670 | 97 | 406 | 4202 |
| 1987 | 771 | 225 | 416 | 1514 | 71 | 507 | 97 | 373 | 3974 |
| 1988 | 853 | 246 | 416 | 1519 | 71 | 607 | 97 | 446 | 4256 |
| 1989 | 598 | 182 | 416 | 1550 | 71 | 522 | 97 | 384 | 3821 |
| 1990 | 770 | 237 | 322 | 1400 | 71 | 416 | 97 | 379 | 3692 |
| 1991 | 625 | 339 | 384 | 1804 | 71 | 627 | 97 | 368 | 4315 |
| 1992 | 714 | 378 | 452 | 1675 | 71 | 803 | 97 | 475 | 4667 |
| 1993 | 356 | 337 | 445 | 1713 | 71 | 855 | 97 | 349 | 4224 |
| 1994 | 462 | 223 | 355 | 1651 | 71 | 821 | 97 | 424 | 4103 |
| 1995 | 376 | 275 | 581 | 1612 | 71 | 715 | 97 | 395 | 4122 |
| 1996 | 660 | 342 | 531 | 1590 | 71 | 682 | 97 | 394 | 4368 |
| 1997 | 786 | 263 | 449 | 1727 | 71 | 574 | 97 | 489 | 4455 |
| 1998 | 746 | 272 | 406 | 1776 | 71 | 613 | 97 | 500 | 4481 |
| 1999 | 843 | 216 | 519 | 1674 | 71 | 586 | 97 | 429 | 4434 |
| 2000 | 1022 | 238 | 495 | 1839 | 71 | 467 | 97 | 467 | 4696 |
| 2001 | 745 | 297 | 777 | 1731 | 71 | 682 | 97 | 379 | 4779 |
| 2002 | 688 | 336 | 779 | 1821 | 71 | 791 | 97 | 335 | 4918 |
| 2003 | 732 | 288 | 545 | 1809 | 71 | 870 | 97 | 244 | 4656 |
| 2004 | 722 | 275 | 533 | 1716 | 71 | 935 | 97 | 247 | 4596 |
| 2005 | 642 | 131 | 492 | 1760 | 71 | 763 | 97 | 402 | 4358 |
| 2006 | 683 | 183 | 473 | 2090 | 71 | 626 | 97 | 399 | 4621 |
| 2007 | 827 | 155 | 443 | 1814 | 71 | 502 | 97 | 286 | 4195 |
| 2008 | 613 | 175 | 374 | 1743 | 71 | 552 | 97 | 247 | 3872 |
| 2009 | 752 | 185 | 429 | 1970 | 71 | 742 | 97 | 268 | 4514 |
| 2010 | 399 | 129 | 611 | 1916 | 71 | 781 | 97 | 224 | 4228 |
| 2011 | 964 | 135 | 464 | 1742 | 71 | 747 | 97 | 236 | 4455 |
| 2012 | 585 | 190 | 578 | 1721 | 71 | 607 | 97 | 213 | 4062 |
| 2013 | 730 | 173 | 611 | 1907 | 71 | 538 | 97 | 186 | 4314 |
| 2014 | 770 | 108 | 578 | 1829 | 71 | 416 | 97 | 414 | 4284 |
| 2015 | 903 | 106 | 628 | 1939 | 71 | 421 | 97 | 417 | 4583 |
| 2016 | 975 | 152 | 641 | 1811 | 71 | 350 | 97 | 348 | 4445 |
| 2017 | 1116 | 157 | 584 | 1993 | 71 | 484 | 97 | 405 | 4907 |
| 2018 | 745 | 208 | 695 | 1909 | 71 | 454 | 97 | 370 | 4549 |

## APPENDIX 2 CPUE INDICES

Table A2: Standardised CPUE indices included in the stock assessment model.

| Year | TAR3-BT | TAR3-SN | TAR2BPLE-BT | ENLD-BT |
| :--- | ---: | ---: | ---: | ---: |
|  |  |  |  |  |
| 1990 | 1.080 | 1.283 | 0.838 |  |
| 1991 | 1.213 | 1.248 | 0.747 |  |
| 1992 | 1.299 | 1.394 | 0.761 |  |
| 1993 | 0.793 | 1.107 | 0.813 |  |
| 1994 | 0.815 | 0.767 | 0.806 | 1.309 |
| 1995 | 0.939 | 1.202 | 0.864 | 0.821 |
| 1996 | 1.179 | 1.047 | 0.944 | 1.370 |
| 1997 | 1.131 | 1.041 | 1.086 | 1.459 |
| 1998 | 1.160 | 1.072 | 0.995 | 1.155 |
| 1999 | 0.922 | 1.069 | 1.253 | 1.006 |
| 2000 | 1.564 | 0.912 | 1.318 | 0.848 |
| 2001 | 1.208 | 1.030 | 1.455 | 0.926 |
| 2002 | 1.126 | 1.511 | 1.931 | 1.160 |
| 2003 | 0.886 | 1.262 | 1.695 | 1.110 |
| 2004 | 0.706 | 1.092 | 1.562 | 1.073 |
| 2005 | 0.707 | 0.929 | 1.116 | 1.458 |
| 2006 | 0.797 | 0.942 | 0.858 | 1.130 |
| 2007 | 0.982 | 0.752 | 0.739 | 0.849 |
| 2008 | 0.869 | 0.759 | 0.760 | 0.982 |
| 2009 | 0.753 | 0.939 | 0.867 | 1.147 |
| 2010 | 0.656 | 0.829 | 0.929 | 0.990 |
| 2011 | 0.994 | 0.784 | 0.728 | 0.893 |
| 2012 | 0.768 | 1.114 | 0.716 | 0.777 |
| 2013 | 0.869 | 0.808 | 0.843 | 0.856 |
| 2014 | 1.033 | 0.749 | 0.745 | 0.709 |
| 2015 | 1.156 | 0.682 | 0.882 | 0.663 |
| 2016 | 1.110 | 0.840 | 0.866 | 0.714 |
| 2017 | 1.284 | 0.839 | 1.097 | 0.596 |
| 2018 | 1.038 | 1.037 | 0.966 | 0.401 |

## APPENDIX 3 FORWARD PROJECTIONS

Table A3a: Summary of the results of the forward projections. Part 1.

| Model year | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 | 2024 | 2025 | 2026 | 2027 | 2028 | 2029 | 2030 | 2031 | 2032 | 2033 | 2034 | 2035 | 2036 | 2037 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Projection year |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |
| Total catch (model) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 4549 | 3727 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 |
| catch 60\% | 4549 | 3727 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 |
| catch 70\% | 4549 | 3727 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 |
| catch $80 \%$ | 4549 | 3727 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 |
| catch $90 \%$ | 4549 | 3727 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 |
| catch $100 \%$ | 4549 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 |
| $\operatorname{Pr}\left(\mathrm{SByr}>40 \% \mathrm{SB}_{0}\right)$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 0.000 | 0.000 | 0.000 | 0.000 | 0.006 | 0.051 | 0.095 | 0.215 | 0.309 | 0.417 | 0.507 | 0.594 | 0.653 | 0.718 | 0.766 | 0.824 | 0.860 | 0.886 | 0.908 | 0.926 |
| catch 60\% | 0.000 | 0.000 | 0.000 | 0.000 | 0.004 | 0.038 | 0.063 | 0.139 | 0.218 | 0.308 | 0.378 | 0.452 | 0.529 | 0.588 | 0.635 | 0.675 | 0.723 | 0.766 | 0.816 | 0.839 |
| catch 70\% | 0.000 | 0.000 | 0.000 | 0.000 | 0.004 | 0.024 | 0.035 | 0.088 | 0.151 | 0.201 | 0.268 | 0.322 | 0.379 | 0.433 | 0.489 | 0.534 | 0.582 | 0.614 | 0.668 | 0.688 |
| catch 80\% | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 0.015 | 0.027 | 0.053 | 0.097 | 0.142 | 0.190 | 0.229 | 0.256 | 0.296 | 0.339 | 0.380 | 0.431 | 0.460 | 0.490 | 0.528 |
| catch $90 \%$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 0.011 | 0.017 | 0.036 | 0.064 | 0.093 | 0.125 | 0.156 | 0.177 | 0.201 | 0.224 | 0.258 | 0.277 | 0.306 | 0.328 | 0.343 |
| catch 100\% | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.009 | 0.009 | 0.019 | 0.037 | 0.065 | 0.081 | 0.101 | 0.119 | 0.129 | 0.138 | 0.154 | 0.169 | 0.195 | 0.218 | 0.245 |
| $\operatorname{Pr}\left(\mathrm{SByr}>35 \% \mathrm{SB}_{0}\right)$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 0.000 | 0.000 | 0.000 | 0.006 | 0.037 | 0.163 | 0.265 | 0.396 | 0.529 | 0.617 | 0.700 | 0.758 | 0.817 | 0.860 | 0.895 | 0.920 | 0.948 | 0.961 | 0.973 | 0.980 |
| catch 60\% | 0.000 | 0.000 | 0.000 | 0.003 | 0.027 | 0.106 | 0.184 | 0.296 | 0.405 | 0.487 | 0.580 | 0.654 | 0.695 | 0.750 | 0.797 | 0.832 | 0.872 | 0.889 | 0.908 | 0.930 |
| catch 70\% | 0.000 | 0.000 | 0.000 | 0.001 | 0.015 | 0.072 | 0.114 | 0.219 | 0.288 | 0.377 | 0.438 | 0.515 | 0.570 | 0.615 | 0.656 | 0.694 | 0.737 | 0.776 | 0.814 | 0.842 |
| catch 80\% | 0.000 | 0.000 | 0.000 | 0.000 | 0.012 | 0.048 | 0.081 | 0.139 | 0.201 | 0.271 | 0.323 | 0.362 | 0.412 | 0.462 | 0.508 | 0.544 | 0.586 | 0.612 | 0.657 | 0.677 |
| catch $90 \%$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.007 | 0.033 | 0.042 | 0.092 | 0.136 | 0.174 | 0.230 | 0.259 | 0.288 | 0.313 | 0.348 | 0.394 | 0.428 | 0.454 | 0.483 | 0.510 |
| catch 100\% | 0.000 | 0.000 | 0.000 | 0.000 | 0.005 | 0.021 | 0.031 | 0.054 | 0.090 | 0.123 | 0.153 | 0.174 | 0.193 | 0.219 | 0.235 | 0.261 | 0.272 | 0.288 | 0.311 | 0.330 |
| $\operatorname{Pr}\left(\mathrm{SByr}>20 \% \mathrm{SB}_{0}\right)$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 0.049 | 0.074 | 0.307 | 0.728 | 0.876 | 0.949 | 0.976 | 0.982 | 0.984 | 0.990 | 0.991 | 0.995 | 0.998 | 0.998 | 0.998 | 0.999 | 0.999 | 0.999 | 0.999 | 1.000 |
| catch 60\% | 0.049 | 0.074 | 0.275 | 0.666 | 0.812 | 0.902 | 0.940 | 0.953 | 0.963 | 0.976 | 0.985 | 0.987 | 0.991 | 0.992 | 0.996 | 0.997 | 0.998 | 0.998 | 0.997 | 0.999 |
| catch 70\% | 0.049 | 0.074 | 0.243 | 0.567 | 0.723 | 0.833 | 0.887 | 0.910 | 0.919 | 0.930 | 0.940 | 0.953 | 0.968 | 0.970 | 0.977 | 0.983 | 0.989 | 0.991 | 0.993 | 0.994 |
| catch 80\% | 0.049 | 0.074 | 0.213 | 0.476 | 0.623 | 0.725 | 0.788 | 0.809 | 0.836 | 0.858 | 0.876 | 0.885 | 0.903 | 0.919 | 0.939 | 0.944 | 0.953 | 0.957 | 0.962 | 0.973 |
| catch 90\% | 0.049 | 0.074 | 0.177 | 0.395 | 0.500 | 0.616 | 0.665 | 0.691 | 0.717 | 0.731 | 0.760 | 0.779 | 0.791 | 0.815 | 0.840 | 0.852 | 0.871 | 0.874 | 0.886 | 0.896 |
| catch $100 \%$ | 0.049 | 0.074 | 0.133 | 0.318 | 0.411 | 0.493 | 0.520 | 0.553 | 0.581 | 0.601 | 0.624 | 0.639 | 0.658 | 0.677 | 0.684 | 0.697 | 0.717 | 0.721 | 0.749 | 0.761 |


| Model year | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 | 2024 | 2025 | 2026 | 2027 | 2028 | 2029 | 2030 | 2031 | 2032 | 2033 | 2034 | 2035 | 2036 | 2037 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Projection year |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |
| $\operatorname{Pr}\left(\mathrm{SByr}>10 \% \mathrm{SB}_{0}\right.$ ) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 0.991 | 0.983 | 0.997 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| catch 60\% | 0.991 | 0.983 | 0.994 | 0.999 | 0.999 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| catch 70\% | 0.991 | 0.983 | 0.990 | 0.998 | 0.999 | 0.999 | 0.999 | 0.999 | 0.998 | 0.999 | 0.999 | 0.998 | 0.998 | 0.998 | 0.998 | 0.998 | 0.998 | 0.998 | 0.999 | 0.999 |
| catch $80 \%$ | 0.991 | 0.983 | 0.988 | 0.998 | 0.997 | 0.997 | 0.997 | 0.994 | 0.995 | 0.993 | 0.991 | 0.992 | 0.992 | 0.995 | 0.996 | 0.996 | 0.997 | 0.998 | 0.997 | 0.997 |
| catch $90 \%$ | 0.991 | 0.983 | 0.983 | 0.996 | 0.996 | 0.995 | 0.990 | 0.982 | 0.976 | 0.977 | 0.978 | 0.979 | 0.982 | 0.978 | 0.979 | 0.977 | 0.979 | 0.981 | 0.983 | 0.984 |
| catch $100 \%$ | 0.991 | 0.983 | 0.978 | 0.991 | 0.987 | 0.980 | 0.969 | 0.952 | 0.948 | 0.937 | 0.931 | 0.933 | 0.928 | 0.927 | 0.929 | 0.925 | 0.924 | 0.923 | 0.924 | 0.925 |
| Median SByr/SB ${ }_{0}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 0.158 | 0.160 | 0.183 | 0.223 | 0.252 | 0.283 | 0.307 | 0.331 | 0.353 | 0.378 | 0.402 | 0.424 | 0.445 | 0.463 | 0.482 | 0.498 | 0.517 | 0.527 | 0.542 | 0.559 |
| catch 60\% | 0.158 | 0.160 | 0.179 | 0.214 | 0.239 | 0.266 | 0.287 | 0.307 | 0.327 | 0.348 | 0.369 | 0.389 | 0.407 | 0.423 | 0.441 | 0.456 | 0.473 | 0.482 | 0.496 | 0.510 |
| catch 70\% | 0.158 | 0.160 | 0.175 | 0.206 | 0.226 | 0.250 | 0.266 | 0.283 | 0.300 | 0.317 | 0.336 | 0.353 | 0.370 | 0.383 | 0.398 | 0.412 | 0.426 | 0.435 | 0.446 | 0.460 |
| catch 80\% | 0.158 | 0.160 | 0.171 | 0.198 | 0.213 | 0.233 | 0.245 | 0.259 | 0.272 | 0.286 | 0.302 | 0.315 | 0.329 | 0.341 | 0.352 | 0.365 | 0.377 | 0.386 | 0.396 | 0.409 |
| catch $90 \%$ | 0.158 | 0.160 | 0.168 | 0.189 | 0.200 | 0.216 | 0.224 | 0.233 | 0.245 | 0.256 | 0.268 | 0.276 | 0.288 | 0.297 | 0.304 | 0.315 | 0.326 | 0.334 | 0.343 | 0.354 |
| catch 100\% | 0.158 | 0.160 | 0.164 | 0.181 | 0.188 | 0.199 | 0.203 | 0.208 | 0.216 | 0.224 | 0.231 | 0.237 | 0.246 | 0.250 | 0.255 | 0.264 | 0.274 | 0.278 | 0.288 | 0.296 |

Table A3a: Summary of the results of the forward projections. Part 2.

| Model year | 2038 | 2039 | 2040 | 2041 | 2042 | 2043 | 2044 | 2045 | 2046 | 2047 | 2048 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Projection year | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 |
| Total catch (model) |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 | 1949 |
| catch 60\% | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 | 2304 |
| catch 70\% | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 | 2659 |
| catch $80 \%$ | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 | 3017 |
| catch 90\% | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 | 3373 |
| catch $100 \%$ | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 | 3727 |
| $\operatorname{Pr}\left(\mathrm{SByr}>40 \% \mathrm{SB}_{0}\right)$ |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 0.948 | 0.961 | 0.973 | 0.973 | 0.976 | 0.979 | 0.984 | 0.990 | 0.990 | 0.990 | 0.993 |
| catch 60\% | 0.864 | 0.881 | 0.907 | 0.927 | 0.939 | 0.947 | 0.950 | 0.961 | 0.964 | 0.970 | 0.975 |
| catch 70\% | 0.743 | 0.771 | 0.794 | 0.816 | 0.823 | 0.848 | 0.864 | 0.873 | 0.890 | 0.896 | 0.903 |
| catch $80 \%$ | 0.552 | 0.580 | 0.623 | 0.641 | 0.667 | 0.685 | 0.701 | 0.712 | 0.729 | 0.753 | 0.770 |
| catch $90 \%$ | 0.374 | 0.388 | 0.420 | 0.465 | 0.488 | 0.506 | 0.515 | 0.537 | 0.560 | 0.565 | 0.587 |
| catch 100\% | 0.255 | 0.273 | 0.280 | 0.299 | 0.329 | 0.346 | 0.354 | 0.375 | 0.401 | 0.413 | 0.422 |
| $\operatorname{Pr}\left(\mathrm{SByr}>35 \% \mathrm{SB}_{0}\right)$ |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 0.988 | 0.990 | 0.995 | 0.993 | 0.995 | 0.993 | 0.994 | 0.994 | 0.995 | 0.997 | 0.999 |
| catch 60\% | 0.949 | 0.960 | 0.970 | 0.975 | 0.975 | 0.979 | 0.983 | 0.989 | 0.989 | 0.989 | 0.992 |
| catch 70\% | 0.859 | 0.880 | 0.897 | 0.915 | 0.934 | 0.940 | 0.945 | 0.953 | 0.957 | 0.961 | 0.969 |
| catch $80 \%$ | 0.719 | 0.754 | 0.776 | 0.792 | 0.809 | 0.818 | 0.838 | 0.851 | 0.864 | 0.876 | 0.879 |
| catch $90 \%$ | 0.531 | 0.549 | 0.590 | 0.609 | 0.627 | 0.637 | 0.664 | 0.679 | 0.693 | 0.709 | 0.723 |
| catch 100\% | 0.356 | 0.366 | 0.388 | 0.430 | 0.456 | 0.462 | 0.475 | 0.511 | 0.524 | 0.528 | 0.545 |
| $\operatorname{Pr}\left(\mathrm{SByr}>20 \% \mathrm{SB}_{0}\right.$ ) |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| catch 60\% | 0.999 | 0.999 | 0.999 | 0.999 | 0.999 | 1.000 | 1.000 | 0.999 | 1.000 | 1.000 | 1.000 |
| catch 70\% | 0.996 | 0.996 | 0.996 | 0.996 | 0.997 | 0.997 | 0.997 | 0.996 | 0.997 | 0.997 | 0.999 |
| catch 80\% | 0.976 | 0.980 | 0.986 | 0.989 | 0.987 | 0.987 | 0.990 | 0.990 | 0.990 | 0.993 | 0.996 |
| catch $90 \%$ | 0.906 | 0.918 | 0.930 | 0.938 | 0.949 | 0.949 | 0.951 | 0.946 | 0.958 | 0.959 | 0.962 |
| catch 100\% | 0.784 | 0.799 | 0.806 | 0.815 | 0.824 | 0.829 | 0.843 | 0.855 | 0.857 | 0.864 | 0.874 |
| $\operatorname{Pr}\left(\mathrm{SByr}>10 \% \mathrm{SB}_{0}\right)$ |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| catch 60\% | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| catch 70\% | 0.999 | 0.999 | 0.999 | 0.999 | 0.999 | 0.999 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| catch 80\% | 0.997 | 0.997 | 0.995 | 0.997 | 0.998 | 0.997 | 0.998 | 0.997 | 0.998 | 0.999 | 1.000 |
| catch $90 \%$ | 0.987 | 0.990 | 0.987 | 0.986 | 0.988 | 0.989 | 0.989 | 0.988 | 0.988 | 0.990 | 0.990 |
| catch 100\% | 0.935 | 0.943 | 0.945 | 0.947 | 0.948 | 0.949 | 0.953 | 0.956 | 0.961 | 0.958 | 0.957 |
| Median SByr/SB ${ }_{0}$ |  |  |  |  |  |  |  |  |  |  |  |
| catch 50\% | 0.569 | 0.576 | 0.592 | 0.608 | 0.617 | 0.622 | 0.629 | 0.642 | 0.650 | 0.656 | 0.664 |
| catch 60\% | 0.520 | 0.526 | 0.542 | 0.558 | 0.567 | 0.571 | 0.578 | 0.590 | 0.601 | 0.604 | 0.611 |
| catch 70\% | 0.469 | 0.476 | 0.490 | 0.506 | 0.514 | 0.519 | 0.525 | 0.537 | 0.546 | 0.549 | 0.556 |
| catch 80\% | 0.415 | 0.422 | 0.433 | 0.449 | 0.457 | 0.464 | 0.468 | 0.480 | 0.486 | 0.489 | 0.498 |
| catch $90 \%$ | 0.360 | 0.365 | 0.375 | 0.388 | 0.396 | 0.402 | 0.406 | 0.416 | 0.420 | 0.425 | 0.433 |
| catch 100\% | 0.298 | 0.304 | 0.313 | 0.322 | 0.334 | 0.336 | 0.342 | 0.352 | 0.359 | 0.364 | 0.369 |

## APPENDIX 4 PRELIMINARY EVALUATION OF POTENTIAL REFERENCE POINTS

The Fisheries New Zealand Harvest Strategy Standard (HSS) specifies a default target biomass level for Fishstocks based on the productivity of the species (Ministry of Fisheries 2008 and 2011). Within the HSS framework, tarakihi is defined as a low productivity stock and the corresponding default target biomass level is $40 \%$ of unexploited spawning biomass ( $S B_{0}$ ). A key criterion for defining an appropriate target biomass level is to minimise the risk of the stock declining below a threshold level that may affect the productivity of the stock, by default this threshold level has been established as the "soft limit" of 20\% SBo (Caddy \& Mahon 1995, Ministry of Fisheries 2008 and 2011).

A simulation analysis was conducted to investigate the performance of alternative target biomass levels for the eastern tarakihi stock, following a similar approach used by Francis \& Mace (2005) and Haist (2014). The target biomass was evaluated within the configuration of a Harvest Control Rule (HCR) and the associated Management Procedures (MP) (Hoggarth et al. 2006, Geromont \& Butterworth 2015).

## Simulation framework

The simulation analysis was implemented in the $R$ software environment. The stock dynamics of the base Stock Synthesis assessment model (updated in 2018) were replicated in the customised $R$ code. The simulator was initialised with the population age structure (numbers-at-age) from the terminal year of the stock assessment model (end of 2017) and conducted a forward projection of the population, incorporating the key parameters from the assessment model (primarily selectivity parameters) to extract fishery specific catches. Each iteration of the simulator utilised the outputs (numbers-at-age and parameter estimates) from an individual MCMC draw from the assessment model. For each MP scenario evaluated, 1000 iterations of the forward projection were conducted (corresponding to the 1000 MCMC samples).

Annual recruitments in the projection period were modelled using three different approaches: 1) individual recruitment deviates were resampled from the last 20 years of the assessment model and applied to the SRR (steepness $h=0.9$ and $R_{0}$ parameter values); 2) individual recruitment deviates were sampled from the assumed distribution of recruitment deviates in the assessment model (sigmaR $=0.6$ ) and applied to the SRR; or 3 ) numbers of recruits were resampled from the last 20 years of the assessment model.

A preliminary set of simulations was also conducted that included autocorrelation in recruitment, as well as simulations incorporating uncertainty in both the natural mortality and SRR steepness parameters.

The forward projections were conducted for a 200 year period. The long evaluation period enabled sufficient time for the model biomass to stabilise at a level that corresponded to the specific HCR being evaluated. Annual catches in the projection period were based on the fishery specific catches that were included in the assessment model for 2017 (termed CatchBase, representing a total catch of 4442 t ). The annual catches in the first year of the projection (2018) were set at the CatchBase level. Annual catches in the five subsequent years (2019-2022) were held constant at 75\% of CatchBase to replicate the level of catches required to initiate the rebuilding of the stock (from $0.16 \% S B_{0}$ in 2017).

Catches during the remainder of the projection period were determined based on the specific Harvest Control Rule (HCR) included in the Management Procedure (MP). The set of HCRs that were evaluated all had the same generic structure, with the total annual catch determined based on the status of the stock relative to the target biomass level (expressed as a proportion of $S B_{0}$ ) (Figure A1). The general structure of the HCR maintains catches at a constant (specified) level when the stock is close to the specified target biomass level (target biomass $\pm 0.05$ threshold) and reduces catches when the stock is below the target biomass range. Annual catches are set to zero when the biomass is below the hard limit $\left(10 \% S B_{0}\right)$. When the biomass is above the upper threshold of the target biomass (target biomass + 0.05 ), the HCR increases catches above the specified base level in a linear manner (Figure A1).

The set of HCRs evaluated included seven base levels of catch that were multiples of the CatchBase (catch scalars of $70 \%, 80 \%, 90 \%, 100 \%, 110 \%, 120 \%$ and $130 \%$ ) at three alternate levels of target
biomass ( $30 \%, 35 \%$ and $40 \% S B_{0}$ ) (Figure A1). The base level of catch for each HCR was the catch taken when the stock was at the respective target.


Figure A1. The range of harvest control rules included in the management procedure evaluation. The grey lines represent the total level of catch associated with multiples of the base level of catch (catch scalars of $\mathbf{7 0 \%}, \mathbf{8 0 \%}, \mathbf{9 0 \%}, \mathbf{1 0 0 \%}, \mathbf{1 1 0 \%}, \mathbf{1 2 0 \%}$ and $\mathbf{1 3 0 \%}$ ) for three alternate levels of notional target biomass ( $\mathbf{3 0 \%}, \mathbf{3 5 \%}$ and $40 \% S B_{0}$ ).

For each projection, the status of the stock was determined at 5-year intervals, approximating the current time frame for conducting full stock assessments of the eastern tarakihi stock. Stock status (SBratio year) was simply determined as $S B_{\text {year }} / S B_{0}+$ Error. The Error component was assumed to be normally distributed with a coefficient of variation of $20 \%$. The magnitude of the error was based on the precision of the estimate of the stock status in the terminal year of the stock assessment (CV 15\%).

The SBratio $_{\text {year }}$ was applied to the specific HCR to determine the corresponding level of catch (CatchMP ${ }_{\text {year }}$ ). This was the basis for determining the level of annual catches in the subsequent 5 -year period (until the next assessment). However, if the difference in the resultant level of catch (CatchMP year) was less than $10 \%$ of the current level of catch (Catch ${ }_{\text {year }}$ ) there was no change in the annual catches implemented for the subsequent period. Further, the maximum change in catch that was implemented (for the next 5 years) was limited to $\pm 20 \%$ of the current level of catch.
Each HCR was evaluated using the set of 1000 MCMC samples. Performance metrics were determined for the projection period, following the exclusion of the first 25 years which included the initial rebuild period. The primary performance metrics were:
i. The proportion of years the biomass is below the $20 \% S B_{0}$ soft limit (all years, all simulations combined) (RiskSB2O).
ii. The proportion of years the biomass is below the $10 \% S B_{0}$ soft limit (all years, all simulations combined) (RiskSB10).
iii. The median spawning biomass ( $\% S B_{0}$ ) during the projection period.
iv. The average annual catch during the projection period.
v. The standard deviation of the annual catches during the projection period.

The performance of individual HCRs was assessed based on the risk criteria specified in the HSS (Ministry of Fisheries 2008). The risk criteria are defined as "the probability of breaching the soft limit $\left(20 \% S B_{0}\right)$ does not exceed $10 \%$ and the probability of breaching the hard limit ( $10 \% S B_{0}$ ) does not exceed $2 \%$ " (criterion 1) or "no more than a 5\% probability of breaching the soft limit" (criterion 2) (Ministry of Fisheries 2008).
Standardised CPUE indices from four main fisheries represent a key input in the eastern tarakihi stock assessment. These are primarily derived from target fisheries and there is potential for the CPUE indices to deviate from the assumed proportional relationship with stock abundance. A set of simulations were
conducted that assumed a degree of hyperstability in the estimates of stock status, representing a proxy for the CPUE indices. This was implemented by including the power term (X) in the determination of the current stock status; i.e. SBratio year $=\left(S B_{\text {year }} / S B_{0}\right)^{\mathrm{X}}+$ Error where $\mathrm{X}=0.9$. This introduced a $10-$ $15 \%$ positive bias in the estimate of stock status at $25-40 \% S B_{0}$.

Additional simulations were conducted that incorporated an additional source of assessment bias that was auto correlated with periodic shifts in the magnitude of the bias. This was implemented following Haist (2014).

## Results

Initial simulations revealed that there was no appreciable difference in the performance of the MPs for the three different formulations of recruitment during the projection period. The final set of simulations used the first recruitment option; i.e. SRR with resampled recruitment deviates from the last 20 years of the assessment.

The base set of simulations revealed that the overall level of biomass during the projection period was more strongly influenced by the catch scalar (and hence magnitude of the catch) than the notional target biomass. For example, for the HCR scenarios with a notional target biomass of $35 \% S B_{0}$, the median level of biomass ranged from $41.5 \% S B_{0}$ for a $70 \%$ catch scalar to $28.5 \% S B_{0}$ for a $130 \%$ catch scalar (Table A4 and Figure A2) . The scenario that realised a level of biomass equivalent to the notional target of $35 \% S B_{0}$ was the HCR with the $90 \%$ catch scalar $\left(35.2 \% S B_{0}\right)$. This scenario produced the level of catch that corresponded to the level of equilibrium yield at the specific target biomass (Table A4).

The range of catch based HCRs were adopted in this study as they more explicitly represent the TACC based management regime. The mismatches between the notional target biomass and the realised level of biomass for individual HCR scenarios is not an indication of a failure of the HCR per se. Rather, it is a consequence of using a catch based HCR which will be less flexible than a fishing mortality based HCR. However, it does indicate that the catch-based HCRs do need to be "tuned" if they are required to achieve the notional target biomass level.
For the sets of HCRs with the notional target of $35 \% S B_{0}$, the HCRs with catch scalars of 70-100\% met both risk criteria, while a catch scalar of $110 \%$ violated risk criterion 1 (Risk20<10\% and Risk10< $2 \%$ ) but not risk criterion 2 (Risk20>5\%) (Table A4 and Figure A2). Catch scalars of $120 \%$ and $130 \%$ failed to meet either one or both of the risk criteria. The realised level of average annual catch was very similar for the $90-130 \%$ range of catch scalars, although annual catches were considerably more variable at the upper range. Somewhat less variable and lower average annual catches were achieved for $70 \%$ and $80 \%$ catch scalars (Table A4 and Figure A2).

The sets of HCRs with the lower notional target of $30 \% S B_{0}$ had a higher probability of breaching the risk criteria compared to the HCRs with the notional target of $35 \% S B_{0}$ and the corresponding catch scalar scenarios and, while the overall levels of average catch were similar, the catches were more variable (for $30 \% S B_{0}$ ). Conversely, the sets of HCRs with the higher notional target of $40 \% S B_{0}$ had a lower probability of breaching the risk criteria compared to the corresponding catch scalar scenarios for the notional target of $35 \% S B_{0}$, while overall levels of average catch were slightly lower and less variable (Table A4 and Figure A2).

For the overall set of HCRs, four scenarios met the risk criteria and yielded higher average annual catches (of about 3850 t ), specifically: notional target biomass $30 \% S B_{0}$ and catch scalar $80 \%$, notional target biomass $35 \% S B_{0}$ and catch scalars of $90 \%$ or $100 \%$, and notional target biomass $40 \% S B_{0}$ and catch scalar $110 \%$. Each of these scenarios realised a median biomass level of about $35 \% S B_{0}$ suggesting that maintaining the stock at that level of biomass (i.e. $35 \% S B_{0}$ ) represents a reasonable trade-off between yield and risk. Of these four scenarios, annual catches were the least variable for the HCR with a notional target biomass $35 \% S B_{0}$ and a catch scalar of $90 \%$ (Table A4 and Figure A2).

Table A4. Performance metrics from the 1000 iterations of each Management Procedure evaluated in the base set of simulations. The individual HCRs are characterised by the Notional Target \% SBo and the catch scalar. The performance metrics include the median biomass level (SB/SB0 \%) , the probability of being below $20 \% S B_{0}$ (Risk20) and $10 \% S B_{0}$ (Risk10) and the average annual catch and associated standard deviation. MPs that breach risk criterion 1 are highlighted in red, MPs that breach risk criterion 2 are highlighted in orange.

| MP criteria |  |
| :--- | :---: |
| Notiona <br> l target | Catch scalar <br> $(\%)$ |
|  |  |
| $30 \%$ | $70 \%$ |
| $30 \%$ | $80 \%$ |
| $30 \%$ | $90 \%$ |
| $30 \%$ | $100 \%$ |
| $30 \%$ | $110 \%$ |
| $30 \%$ | $120 \%$ |
| $30 \%$ | $130 \%$ |
|  |  |
| $35 \%$ | $70 \%$ |
| $35 \%$ | $80 \%$ |
| $35 \%$ | $90 \%$ |
| $35 \%$ | $100 \%$ |
| $35 \%$ | $110 \%$ |
| $35 \%$ | $120 \%$ |
| $35 \%$ | $130 \%$ |
|  |  |
| $40 \%$ | $70 \%$ |
| $40 \%$ | $80 \%$ |
| $40 \%$ | $90 \%$ |
| $40 \%$ | $100 \%$ |
| $40 \%$ | $110 \%$ |
| $40 \%$ | $120 \%$ |
| $40 \%$ | $130 \%$ |
|  |  |


|  |  |  | Performance metrics |  |
| ---: | ---: | ---: | ---: | ---: |
|  |  |  |  | Catch (t) |
|  |  |  |  | Mean | St.dev






| * 30\%SB0catch70\% | - 30\%SB0catch110\% | * 35\%SB0catch80\% | - 35\%SB0catch120\% | 40\%SB0catch90\% | - 40\%SB0catch130\% |
| :---: | :---: | :---: | :---: | :---: | :---: |
| - 30\%SB0catch80\% | - 30\%SB0catch120\% | - 35\%SB0catch90\% | - 35\%SB0catch130\% | - 40\%SB0catch100\% |  |
| - 30\%SB0catch90\% | - 30\%SB0catch130\% | - 35\%SB0catch100\% | - 40\%SB0catch70\% | - 40\%SB0catch110\% |  |
| - 30\%SB0catch100\% | - 35\%SB0catch70\% | - 35\%SB0catch110\% | - 40\%SB0catch80\% | - 40\%SB0catch120\% |  |

Figure A2: Performance metrics relative to the realised level of biomass ( $S B / S B_{0}$ ) from the 1000 iterations of each Management Procedure evaluated in the base set of simulations. The individual HCRs are characterised by the Notional Target \% SBO (colours) and the catch scalar (symbols). The risk criteria associated with the probability of breaching the soft limit (top left panel) and hard limit (top right panel) are specified by the horizontal red lines.

Additional simulations were conducted with additional autocorrelation in assessment bias (Table A5) and hyperstability in the CPUE proxy for the assessment (Table A6). Including autocorrelation in assessment bias resulted in a marginal increase in the risk metrics, without appreciably changing the overall risk profile of the range of HCRs evaluated (relative to the base simulations). However, there
was an increase in the variability of the annual catches, while the average levels of catch remained similar to the levels from the base simulations (Table A5).

The introduction of hyperstability in the CPUE proxy also increased the level of risk relative to the base simulations (Table A6). The impact was most pronounced for the set of HCRs with the $30 \% S B_{0}$ notional target biomass level and extended the risk profile to include a lower tier of catch scalars. The variation in annual catches also tended to increase for the lower target biomass options due to the increase in the frequency of the hard limit being breached (resulting in catches being set to zero).
An additional set of simulations was conducted that incorporated autocorrelation ( $R h o=0.60$ ) in the recruitment deviates. This option resulted in a small increase in Risk20 for HCRs scenarios with lower levels of target biomass ( $30 \%$ and $35 \% S B_{0}$ ) and higher catch levels ( $110 \%, 120 \%$ and $130 \%$ scalars). However, there was no evidence of strong autocorrelation in the estimated recruitment deviates from the stock assessment model $\left(\right.$ Recdev $_{\text {yaar }+1}=0.045 *$ Recdev $\left._{\text {year }}\right)$ and, hence, autocorrelation was not incorporated in the final set of simulations.
In addition, incorporating uncertainty in the natural mortality $\operatorname{Normal(mean~}=0.1, \mathrm{sd}=0.025$ ) and SRR steepness $\mathrm{h} \operatorname{Uniform}(0.75,1.00)$ parameters in the assessment model did not appreciably change the risk profile of the range of HCRs evaluated.

## Summary

The simulator incorporates the generalised population dynamics and fishery dynamics of the assessment model and provides the platform to evaluate a range of alternative management options. The HCRs and associated management procedures are an attempt to codify the management response at different levels of stock biomass relative to the target biomass level. The basic structure of the HCRs is considered to be appropriate from a technical perspective for the purpose of evaluating alternative target biomass levels. The results of the evaluation of alternative HCRs are intended to assist managers in the formulation of an appropriate level of target biomass for the eastern tarakihi stock.
The current study indicates that a target biomass level of about $35 \% S B_{0}$ is sufficiently high to minimise the risk of breaching the soft and hard limits, while maintaining catches at a relatively high level. This is somewhat lower than the current default target biomass level of $40 \% S B_{0}$. These results are predicated on the presumptions regarding the management procedures and the HCRs adopted for the simulations.

The actual management procedure and HCR have not been prescribed for the eastern tarakihi stock. The management strategy for the stock needs to be developed within the framework of a management plan, formulated by fisheries managers in conjunction with the commercial, recreational and customary fishing sectors. The magnitude of the appropriate target biomass will be dependent on key elements of the management strategy, including the proportional distribution of catch amongst fisheries, the frequency of stock assessment process, the accuracy of the assessment and the nature of the management response. To date, there has been limited input from managers into the formulation of the range of HCRs that might be considered appropriate for the management of the tarakihi fishery.

Table A5: Performance metrics from the 1000 iterations of each Management Procedure evaluated in the set of simulations that included autocorrelation in the assessment error. The individual HCRs are characterised by the Notional Target $\% S B_{0}$ and the catch scalar. The performance metrics include the median biomass level (SB/SB $B_{0} \%$ ), the probability of being below $20 \% \operatorname{SB}_{0}$ (Risk20) and $10 \% S B_{0}$ (Risk10) and the average annual catch and associated standard deviation. MPs that breach risk criterion 1 are highlighted in red, MPs that breach risk criterion 2 are highlighted in orange.

| MP criteria |  | Performance metrics |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Notional target | $\begin{array}{r} \text { Catch } \\ \text { scalar }(\%) \end{array}$ | $S B / S B_{0}$ | Risk20 | Risk10 | Catch (t) |  |
|  |  |  |  |  | Mean | St.dev |
| 30\% | 70\% | 38.1\% | 1.1\% | 0.3\% | 3682 | 766 |
| 30\% | 80\% | 34.6\% | 3.1\% | 0.4\% | 3825 | 802 |
| 30\% | 90\% | 31.6\% | 7.7\% | 1.7\% | 3887 | 977 |
| 30\% | 100\% | 29.0\% | 14.2\% | 2.8\% | 3905 | 1190 |
| 30\% | 110\% | 26.9\% | 22.6\% | 5.9\% | 3845 | 1470 |
| 30\% | 120\% | 24.6\% | 33.5\% | 13.8\% | 3590 | 1913 |
| 30\% | 130\% | 22.4\% | 41.9\% | 20.2\% | 3380 | 2203 |
| 35\% | 70\% | 41.6\% | 0.5\% | 0.1\% | 3522 | 723 |
| 35\% | 80\% | 38.2\% | 1.2\% | 0.2\% | 3683 | 761 |
| 35\% | 90\% | 35.4\% | 2.8\% | 0.5\% | 3780 | 851 |
| 35\% | 100\% | 33.2\% | 5.6\% | 0.8\% | 3851 | 961 |
| 35\% | 110\% | 31.3\% | 10.1\% | 1.8\% | 3872 | 1,134 |
| 35\% | 120\% | 29.9\% | 15.2\% | 3.8\% | 3840 | 1,338 |
| 35\% | 130\% | 28.6\% | 21.1\% | 7.2\% | 3738 | 1,600 |
| 40\% | 70\% | 44.9\% | 0.2\% | 0.0\% | 3361 | 698 |
| 40\% | 80\% | 41.8\% | 0.5\% | 0.1\% | 3514 | 738 |
| 40\% | 90\% | 39.1\% | 0.0\% | 0.2\% | 3636 | 807 |
| 40\% | 100\% | 37.1\% | 2.1\% | 0.2\% | 3727 | 870 |
| 40\% | 110\% | 35.4\% | 4.3\% | 0.7\% | 3770 | 992 |
| 40\% | 120\% | 33.9\% | 7.0\% | 1.4\% | 3794 | 1122 |
| 40\% | 130\% | 33.0\% | 10.4\% | 2.8\% | 3758 | 1281 |

Table A6: Performance metrics from the 1000 iterations of each Management Procedure evaluated in the set of simulations that included hyperstability in the CPUE proxy. The individual HCRs are characterised by the Notional Target $\% S B_{0}$ and the catch scalar. The performance metrics include the median biomass level (SB/SB $B_{0} \%$ ), the probability of being below $20 \% \operatorname{SB}_{0}$ (Risk20) and $10 \% S B_{0}$ (Risk10) and the average annual catch and associated standard deviation. MPs that breach risk criterion 1 are highlighted in red, MPs that breach risk criterion 2 are highlighted in orange.

| MP criteria |  | Performance metrics |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Notional | Catch | $S B / S B_{0}$ | Risk20 | Risk10 | Catch (t) |  |
|  |  |  |  |  | Mean | St.dev |
| 30\% | 70\% | 35.3\% | 1.7\% | 0.1\% | 3809 | 717 |
| 30\% | 80\% | 31.6\% | 5.4\% | 0.6\% | 3941 | 766 |
| 30\% | 90\% | 28.2\% | 12.3\% | 1.0\% | 4023 | 848 |
| 30\% | 100\% | 25.6\% | 23.4\% | 2.8\% | 4026 | 1057 |
| 30\% | 110\% | 23.1\% | 35.8\% | 8.3\% | 3882 | 1448 |
| 30\% | 120\% | 20.7\% | 47.2\% | 17.0\% | 3628 | 1876 |
| 30\% | 130\% | 17.0\% | 59.8\% | 32.2\% | 3064 | 2310 |
| 35\% | 70\% | 38.9\% | 0.5\% | 0.0\% | 3659 | 675 |
| 35\% | 80\% | 35.3\% | 1.7\% | 0.1\% | 3813 | 703 |
| 35\% | 90\% | 32.2\% | 4.2\% | 0.2\% | 3929 | 745 |
| 35\% | 100\% | 29.6\% | 9.5\% | 0.6\% | 3994 | 849 |
| 35\% | 110\% | 27.6\% | 17.0\% | 2.2\% | 3991 | 1053 |
| 35\% | 120\% | 25.9\% | 25.1\% | 5.4\% | 3926 | 1315 |
| 35\% | 130\% | 24.1\% | 33.7\% | 11.0\% | 3764 | 1643 |
| 40\% | 70\% | 42.2\% | 0.1\% | 0.0\% | 3498 | 654 |
| 40\% | 80\% | 38.7\% | 0.4\% | 0.0\% | 3669 | 671 |
| 40\% | 90\% | 35.8\% | 0.0\% | 0.2\% | 3790 | 737 |
| 40\% | 100\% | 33.6\% | 3.3\% | 0.2\% | 3874 | 809 |
| 40\% | 110\% | 31.7\% | 6.4\% | 0.3\% | 3936 | 886 |
| 40\% | 120\% | 30.1\% | 11.3\% | 1.5\% | 3946 | 1032 |
| 40\% | 130\% | 29.1\% | 16.0\% | 3.0\% | 3931 | 1186 |

